Traffic-related air pollution in relation to respiratory symptoms, allergic sensitisation and lung function in schoolchildren

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ABSTRACT

Background: Urban air pollution can trigger asthma exacerbations, but the effects of long-term exposure to traffic-related air pollution on lung function or onset of airway disease and allergic sensitisation in children is less clear.

Methods: All 2107 children aged 9–14 years from 40 schools in Rome in 2000–1 were included in a cross-sectional survey. Respiratory symptoms were assessed on 1760 children by parental questionnaires (response rate 83.5%). Allergic sensitisation was measured by skin prick tests and lung function was measured by spirometry on 1359 children (77.2%). Three indicators of traffic-related air pollution exposure were assessed: self-reported heavy traffic outside the child’s home; the measured distance between the child’s home and busy roads; and the residential nitrogen dioxide (NO2) levels estimated by a land use regression model (R² = 0.68).

Results: There was a strong association between estimated NO2 exposure per 10 μg/m³ and lung function, especially expiratory flows, in linear regression models adjusted for age, gender, height and weight: −0.62% (95% CI −1.05 to −0.19) for forced expiratory volume in 1 s as a percentage of forced vital capacity, −62 ml/s (95% CI −102 to −21) for forced expiratory flow between 25% and 75% of forced vital capacity and −85 ml/s (95% CI −135 to −35) for peak expiratory flow. The other two exposure indicators showed similar but weaker associations. The associations appeared stronger in older children, in children of high socioeconomic status and in those exposed to parental smoking. Although lifetime asthma was not an effect modifier, there was a suggestion of a larger effect on lung function in subjects with a positive prick test. Multiple logistic regression models did not suggest a consistent association between traffic-related air pollution exposure and prevalence of respiratory symptoms or allergic sensitisation.

Conclusion: The results of this study suggest that residential traffic-related air pollution exposure is associated with reduced expiratory flows in schoolchildren.

Numerous studies have reported adverse effects of urban air pollution on various aspects of respiratory health in children. Traffic-related air pollution has been associated with asthma exacerbations, doctor-diagnosed asthma and increased prevalence of asthma symptoms. There are also studies demonstrating an association between traffic pollution and sensitisation to pollen and outdoor allergens, as well as increased levels of total and specific IgE.

Data from cross-sectional studies have also suggested that ambient air pollution can affect lung function in 10-year-old children and recent longitudinal cohort data have shown that pollution from traffic has adverse effects on children’s lung development between 10 and 18 years of age. However, other studies have indicated no association between air pollution indicators and respiratory symptoms or lung function.

Previous studies have assessed long-term exposure to traffic air pollution by self-reported traffic intensity, distance to busy roads or estimated pollution levels by dispersion or regression modelling, but no study has yet evaluated in a comparative way all of these indicators of traffic-related air pollution exposure. There is also a lack of studies that objectively assess both exposure and disease, and few investigations in Europe have used population data in an urban area with large contrasts in air pollution concentrations. In addition, there are few studies on traffic-related air pollution as a causative agent in the initial disease development in young school children. Even if both epidemiological and experimental studies have proposed plausible biological explanations of these effects, there is a great need for more scientific evidence, especially in view of recent cohort data from California linking air pollution to impaired lung development.

The aim of this study was to investigate the association between traffic-related air pollution exposure and objective measurements of lung function by spirometry and atopy by skin prick testing and parental reports of respiratory symptoms in a population-based study of children in Rome, Italy. Three different indicators of traffic pollution were used: self-reported traffic level, distance from busy roads measured by a geographic information system and the residential levels of nitrogen dioxide (NO2) estimated by a land use regression model.

METHODS

Study subjects and assessment of respiratory health

The children in this study were enrolled for the Italian part of the second phase of the International Study of Asthma and Allergies in Childhood (ISoAAC phase 2). Selection of participants and data collection were therefore performed according to the ISoAAC methodology as described elsewhere. Children were selected from public and private primary schools in Rome (and in the small nearby municipality of Fiumicino). Parents of all children attending the fifth grade during the 2000–1 school year (n = 2107) from each school.

(n = 40) were contacted. Of these, 1760 (83.5%) parents of children between 9 and 14 years old responded to the Italian version of the ISAAC II questionnaire.

The standardised questionnaire was used to ascertain lifetime prevalence of asthma, wheezy, hay fever and eczema as well as the child’s respiratory health during the past 12 months including wheezing, cough, phlegm and rhinitis.21 On the basis of asthma and wheezing onset and the presence of wheezing in the 12 months before the survey (current wheezing) from questionnaire data, children were divided into four mutually exclusive groups: (1) those with asthma or wheezing onset in the first 2 years of life and no current wheezing (early transient wheezers); (2) those with asthma or wheezing onset in the first 2 years of life who still wheezed at the time of the survey (persistent wheezers); (3) those with asthma or wheezing onset from 3 years onwards who still wheezed at the time of the survey (late onset wheezers); (4) those with no wheezing nor asthma in the first two years of life and no current wheezing (control group).

The Italian questionnaire also included certain factors known to be related with asthma and allergies, in particular those that occurred during pregnancy and the postnatal period. In addition, the questionnaire also included items on the intensity of traffic and the frequency of heavy traffic outside the child’s home. The parents or another person responsible for the child’s care completed the questionnaire.

The parents of each child were informed of the study and provided their consent regarding the health examination, including prick tests and spirometry at school. Informed consent was given by the parents of 1359 children (77.2%). Prick tests were performed to provide an objective measure of atopic sensitisation to Dermatophagoides pteronyssinus, Dermatophagoides farinace, cat fur, mixed grass pollen, mixed tree pollen and Alternaria tenuis using histamine (10 mg/ml) and a saline solution as positive and negative controls. In addition, Olea and Parietaria species were included. All extracts and control solutions were obtained from ALK Laboratories (Horsholm, Denmark). Subjects with one or more positive result were considered atopic, determined by a weal size >3 mm² from markers on mm² paper transferred by tape 15 min after exposure. Height, weight and baseline pulmonary function were measured using a portable spirometer (Biomedin-Padua, Italy).21 The spirometric values for the best blow of at least three attempts were used, according to American Thoracic Society recommendations.22 Children with any respiratory infection during the week before the lung function test were excluded. Lung function parameters measured included forced vital capacity (FVC, ml), forced expiratory volume in 1 s (FEV₁, ml), forced expiratory flow between 25% and 75% of FVC (FEF25–75, ml/s) and peak expiratory flow (PEF, ml/s). The best FVC and best FEV₁ were recorded, whereas FEF25–75 and PEF were derived from the best curve, defined as the greatest sum of FVC and FEV₁. In addition, FEV₁ as a percentage of FVC was calculated.

**Environmental exposure**

Different indicators of residential traffic-related air pollution were assessed for each child: self-reported traffic intensity and frequency of heavy traffic in the area of residence, the distance from home to busy roads and the estimated NO₂ concentrations outside the child’s home. Parents of 1752 children (98.4%) rated traffic intensity as none, little, moderate and heavy, and frequency of heavy traffic as rarely, sometimes or often. A geographical information system (ArcGis 9.1, ESRI, Redlands, California, USA) was used to geocode the residence of each child at the time of the survey. The Municipal Office of Rome provided information on traffic counts that was used to identify roads with dense traffic. We then estimated the distance from these busy roads for the addresses of 1530 children (86.9%) and categorised the distances as >200 m, 100–200 m, 50–100 m, 20–50 m and <20 m (alternatively as more or less than 100 m).

The NO₂ concentration in the census block of residence (a small area of about 500 inhabitants per census block) was estimated for each child by a land use regression model described in detail elsewhere.24 This model was built by assessing the association between variables describing land use and traffic information and NO₂ concentrations measured at 68 sites in Rome. Palme’s tubes measured outdoor pollution simultaneously over three 7-day periods in 1995 and 1996. The most important variables predicting measured NO₂ were the area of the city, distance from busy roads, size of the census block, inverse population density and altitude, resulting in a determination coefficient (R²) of 0.686. A hybrid version of that model was used for the present study where altitude and distance from busy roads were replaced by emission data of benzene obtained from the Municipality of Rome, resulting in an R² of 0.690. The correlation between observed and predicted NO₂ concentrations from the land use regression model was 0.77. Census blocks with extreme values of input data for the prediction model (ie, inverse population and size of the census block) were excluded to avoid extrapolation of predicted NO₂ beyond a reasonable range, resulting in a successful model estimation of residential NO₂ concentrations for 1578 (89.7%) children.

**Potential confounders and effect modifiers**

Several potential confounders of the association between traffic-related air pollution variables and respiratory health were a priori chosen from the questionnaire or from anthropometric measurement: gender, age, height, weight, body mass index (BMI), father’s and mother’s education (less than high school, high school or above), parental smoking habit (current, former, never smoker), mother smoking during pregnancy (yes, no), parental history of asthma (yes, no), pets during the first year of life (yes, no), keeping a cat or a dog ever (yes, no), day care attendance (yes, no), indoor humidity or signs of mould in the child’s bedroom (yes, no), number of siblings and the floor of the subject’s home (ground floor, first–second floor, third–fourth floor, fourth floor or above). We did not take into account gas cooking because 1725 out of 1760 children (95%) lived in a household using it.

We were also a priori interested in effect modification, in particular whether the air pollution effects differed by gender, age (9–10, 11–14 years), maternal education, current parental smoking as indicator of environmental tobacco smoke (ETS) exposure (no, yes), individual BMI (below and above the 80th percentile), antioxidant intake (fresh fruit consumption <3 and ≥3 times/week), reported lifetime asthma (no, yes) and atopic status (negative and positive prick test).

**Statistical analyses**

Spearman’s rank correlation coefficient was used as a non-parametric measure of correlation of the exposure variables. The associations between variables indicating air pollution exposure and lung function parameters were analysed by linear regression. The final multiple linear regression models included only covariates that had a notable impact (10% change) on the
traffic-related air pollution effect estimates (ie, age, gender, height and weight). The results are presented as regression coefficients (beta) and 95% confidence intervals (95% CI) representing a unit change in the parameter for a 10 µg/m³ increase in NO₂. For self-reported categorical variables of exposure to traffic and for categories of distance to busy roads, we estimated the change in lung function for each category compared with the first one and the coefficient across the categories to evaluate the linear trend.

The associations between air pollution exposure and respiratory symptoms or atopic sensitisation were analysed by logistic regression and the results are presented as odds ratios (OR) and 95% confidence intervals (95% CI). All logistic regression models included covariates that influenced the effect of air pollution by 10% or more (ie, age, gender, BMI, father's education, maternal smoking, indoor humidity or signs of mould in the child’s bedroom, and the floor of the subject’s home). The logistic regression analyses were made with an exposure variable representing an interquartile range increase in the NO₂ level (14.2 µg/m³) or dichotomous variables of traffic intensity (no–little vs moderate–heavy), heavy traffic (rarely vs sometimes–often) and distance to roads (more or less than 100 m).

Sensitivity analyses were conducted using alternative cut-off points of traffic intensity, heavy traffic and distance to busy roads (at 100, 150 or 200 m).

Effect modification of the associations between air pollution and respiratory and allergic outcomes was tested by introducing interaction terms in the linear and logistic regression models between NO₂ and the a priori considered variables. All statistical analyses were performed with Stata SE Version 9 (Stata, Texas, USA).

RESULTS
Table 1 shows the general characteristics of the children participating in the study, separately for those who provided clinical and lung function data and for those who did not. The overall prevalence of ever wheezing, asthma and symptoms of rhinitis were 15.4%, 12.6% and 9.1%, respectively. In general, a larger proportion of children who provided clinical data suffered from respiratory symptoms than those who were not clinically tested. Moreover, most of the risk factors related to allergic and airway diseases were more common in the clinically tested group.
Environmental exposure

Figure 1 Predicted nitrogen dioxide (NO2) exposure according to self-reported residential traffic intensity and distance from busy roads of children in Rome 2000–1. The box includes the 25th (lower hinge) to the 75th (upper hinge) percentile and the median (middle bar), the whiskers represent 1.5 times the interquartile range and the circles are outliers.

Table 2 gives descriptive data of the variables collected at the health examination, including anthropometric measures, lung function parameters and prick test positivity. Almost 32% showed signs of atopic sensitisation to at least one of the tested allergens. The most frequent was the positivity to D pteronyssinus (18.7%). In general, boys were more likely than girls to have a positive skin prick test, but the differences were small and not statistically significant (data not shown).

The overall mean individual estimated NO2 exposure was 45 µg/m3 (range 21–63 µg/m3). The correlation between estimated NO2 exposure and self-reported intensity of traffic outside the child’s home was moderate (rxy = 0.57). The mean (and median) estimated air pollution exposure increased with each category of self-reported traffic intensity, from 36.58 µg/m3 for those reporting no traffic to 42.13 µg/m3 for those with little traffic, 47.56 µg/m3 for those with moderate traffic, and up to 51.71 µg/m3 for those reporting intense traffic at home (fig 1). The number of children in these categories was 352, 484, 599 and 297, respectively. The self-reported frequency of heavy traffic was less correlated with estimated NO2 (rxy = 0.21) and demonstrated a similar but less pronounced pattern with mean levels from 43.07 µg/m3 for those who never or rarely experience heavy traffic to 46.28 µg/m3 for those with moderate heavy traffic (a couple of times each day) and 47.74 µg/m3 for those who have trucks passing frequently.

The number of children in these categories was 962, 563 and 207, respectively. The estimated NO2 level did not change substantially up to 100 m away from busy roads, but appeared to decrease with greater distances (fig 1). The mean NO2 level was 50.87 µg/m3 within 20 m, 48.06 µg/m3 20–50 m away, 48.85 µg/m3 at 50–100 m, 47.62 µg/m3 at 100–200 m and 42.65 µg/m3 at 200 m and more, with 98, 53, 128, 219 and

Table 3 Association between different indicators of residential air pollution exposure and lung function measured in children living in Rome 2000–1, presented as regression coefficients (beta) and 95% confidence intervals (95% CI) from linear regression models adjusted for age, gender, height and weight.

<table>
<thead>
<tr>
<th>No of subjects</th>
<th>FVC (ml)</th>
<th>FEV1 (ml)</th>
<th>FEV1/FVC (%)</th>
<th>FEF25–75 (ml/s)</th>
<th>PEF(ml/s)</th>
</tr>
</thead>
<tbody>
<tr>
<td>NO2 (per 10 µg/m3)†</td>
<td>1195</td>
<td>1 (−18 to 21)</td>
<td>−13 (−31 to 5)</td>
<td>−0.62 (−1.05 to −0.19)*</td>
<td>−62 (−102 to −21)*</td>
</tr>
<tr>
<td>Traffic intensity;‡</td>
<td>1320</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>No (reference)</td>
<td>273</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>Little</td>
<td>360</td>
<td>−21 (−69 to 27)</td>
<td>−38 (−81 to 5)</td>
<td>−0.87 (−1.9 to 0.16)</td>
<td>−87 (−184 to 11)</td>
</tr>
<tr>
<td>Moderate</td>
<td>458</td>
<td>−33 (−79 to 13)</td>
<td>−38 (−79 to 3)</td>
<td>−0.37 (−1.35 to 0.61)</td>
<td>−45 (−138 to 48)</td>
</tr>
<tr>
<td>Heavy</td>
<td>229</td>
<td>−13 (−66 to 41)</td>
<td>−49 (−98 to −1)*</td>
<td>−1.63 (−2.78 to −0.47)**</td>
<td>−140 (−250 to −30)*</td>
</tr>
<tr>
<td>Test for trend</td>
<td>−6 (−23 to 10)</td>
<td>−14 (−29 to 0)</td>
<td>−0.4 (−0.72 to 0.02)*</td>
<td>−33 (−66.7 to 0.51)</td>
<td>−76 (−117 to −34)**</td>
</tr>
<tr>
<td>Heavy traffic;‡†</td>
<td>1321</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>Rarely (reference)</td>
<td>743</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>Sometimes</td>
<td>423</td>
<td>33 (−3 to 69)</td>
<td>13 (−20 to 45)</td>
<td>−0.69 (−1.47 to 0.09)</td>
<td>−7 (−82 to 68)</td>
</tr>
<tr>
<td>Often</td>
<td>155</td>
<td>70 (17 to 122)**</td>
<td>34 (−13 to 82)</td>
<td>−1.21 (−2.3 to −0.07)*</td>
<td>−41 (−149 to 67)</td>
</tr>
<tr>
<td>Test for trend</td>
<td>34 (11 to 58)**</td>
<td>16 (−5 to 37)</td>
<td>−0.6 (−1.14 to −0.12)*</td>
<td>−16 (−65 to 32)</td>
<td>−30 (−91 to 30)</td>
</tr>
<tr>
<td>Distance to busy roads (m)$§</td>
<td>1176</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>≥200 (reference)</td>
<td>786</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>100–199</td>
<td>171</td>
<td>46 (−4 to 95)</td>
<td>61 (16 to 106)**</td>
<td>0.83 (−0.27 to 1.93)</td>
<td>97 (−7 to 201)</td>
</tr>
<tr>
<td>50–99</td>
<td>97</td>
<td>11 (−52 to 75)</td>
<td>−2 (−60 to 55)</td>
<td>−0.63 (−2.03 to 0.77)</td>
<td>−73 (−208 to 82)</td>
</tr>
<tr>
<td>20–49</td>
<td>36</td>
<td>13 (−87 to 75)</td>
<td>46 (−45 to 136)</td>
<td>1.35 (−0.87 to 3.56)</td>
<td>80 (−130 to 290)</td>
</tr>
<tr>
<td>≥20</td>
<td>84</td>
<td>76 (8 to 143)*</td>
<td>30 (−31 to 92)</td>
<td>−1.52 (−3.02 to −0.03)*</td>
<td>−22 (−166 to 121)</td>
</tr>
<tr>
<td>Test for trend</td>
<td>16 (1 to 30)*</td>
<td>9 (−4 to 22)</td>
<td>−0.2 (−0.53 to 0.11)</td>
<td>−2 (−33 to 29)</td>
<td>−16 (−54 to 21)</td>
</tr>
</tbody>
</table>

*p<0.05; **p<0.01. †Assessed by a land use regression model. ‡Self-reported traffic intensity. §Objectively estimated on a digital map by the GIS tool ArcGIS. FEF25–76, forced expiratory flow between 25% and 75% of FVC; FEV1, forced expiratory volume in 1 s; FVC, forced vital capacity; PEF, peak expiratory flow.

1032 children, respectively, in each of these categories. The correlation between distance from busy roads and estimated residential NO2 levels was −0.47.

Traffic-related air pollution exposure estimated as the residential NO2 level by a land use regression model was associated with reduced lung function for all spirometric parameters except FVC (table 3). The effect was statistically significant for the FEV1/FVC ratio, FEF25–75 and PEF, with a decrease in lung function measures of 0.62%, 62 ml/s and 23 ml/s per 10 μg/m³ in NO2, respectively. Using self-reported traffic intensity as the relevant exposure presented similar results with a statistically significant negative trend for FEV1/FVC ratio and PEF. The results for heavy traffic exposure and distance from busy roads did not substantially change when the analyses were restricted to participants with complete data on all the exposure indices (n = 1120, data not shown). When other potential confounders such as maternal and paternal smoking, education level and history of asthma were included in the regression model, the estimated decrease in FEF25–75 related to NO2 changed from −62 ml/s (95% CI −102 to −21) to −64 ml/s (95% CI −105 to −23), that is less than 3%, as expected from the strategy chosen in the selection of potential confounders.

In the stratified analyses the effects of NO2 on lung function appeared stronger in girls, older children, those with higher socioeconomic status as indicated by the mother’s education, passive smoke exposure and low fruit consumption, while the other covariates did not suggest any effect modification (table 4). However, the only statistically significant effect modification was age for FEV1 and FEF25–75. The effect modification of gender, maternal education and ETS at home was of borderline significance for FEF25–75. It is to be noted that, while the effect of traffic-related deficit in lung function was present in children with and without asthma, subjects with a positive skin prick test had the largest effect.

In general, none of the respiratory symptoms or any of the skin prick tests appeared to be positively associated with any of the traffic-related air pollution exposure metrics (fig 2), with the exception of moderate or high frequency of heavy traffic exposure which was positively associated with persistent wheezing (OR 1.84, 95% CI 1.01 to 3.10) and any positive skin prick test (OR 1.30, 95% CI 1.00 to 1.70) and a distance of 100 m or less from busy roads presenting a statistically significant positive effect on sensitisation to cat antigens (OR 1.69, 95% CI 1.01 to 2.84). Moreover, there was no effect modification of skin prick test positivity for any of the respiratory symptoms analysed (data not shown). Sensitivity analyses using alternative cut-off points of traffic intensity, heavy traffic and distance to busy roads did not substantially change the results (data not shown).

Table 4  Effect modification of a priori considered variables: regression coefficients (β) and 95% confidence intervals (95% CI) from stratified linear regression models adjusted for age, gender, height and weight

<table>
<thead>
<tr>
<th></th>
<th>FEV1 (ml)</th>
<th>FEF25–75 (ml/s)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>No  Beta (95% CI)</td>
<td>p Value</td>
</tr>
<tr>
<td>Gender</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Boys</td>
<td>711</td>
<td>−4 (−29 to 21)</td>
</tr>
<tr>
<td>Girls</td>
<td>648</td>
<td>−23 (−49 to 2)</td>
</tr>
<tr>
<td>Age (years)</td>
<td></td>
<td></td>
</tr>
<tr>
<td>9–10</td>
<td>1213</td>
<td>−6 (−25 to 12)</td>
</tr>
<tr>
<td>11–14</td>
<td>146</td>
<td>−89 (−154 to −24)</td>
</tr>
<tr>
<td>Maternal education</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Low</td>
<td>538</td>
<td>3 (−27 to 33)</td>
</tr>
<tr>
<td>High</td>
<td>801</td>
<td>−23 (−46 to 0)</td>
</tr>
<tr>
<td>ETS at home</td>
<td></td>
<td></td>
</tr>
<tr>
<td>No</td>
<td>683</td>
<td>−13 (−38 to 11)</td>
</tr>
<tr>
<td>Yes</td>
<td>662</td>
<td>−13 (−39 to 13)</td>
</tr>
<tr>
<td>Fruit intake</td>
<td></td>
<td></td>
</tr>
<tr>
<td>0–2 times/week</td>
<td>377</td>
<td>−5 (−41 to 31)</td>
</tr>
<tr>
<td>&gt;2 times/week</td>
<td>962</td>
<td>−15 (−36 to 5)</td>
</tr>
<tr>
<td>BMI</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Low–normal</td>
<td>1489</td>
<td>−12 (−34 to 3)</td>
</tr>
<tr>
<td>High</td>
<td>271</td>
<td>16 (−57 to 89)</td>
</tr>
<tr>
<td>Asthma</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Never</td>
<td>1152</td>
<td>−16 (−36 to 4)</td>
</tr>
<tr>
<td>Ever</td>
<td>194</td>
<td>−5 (−46 to 37)</td>
</tr>
<tr>
<td>Positivity to prick test</td>
<td></td>
<td></td>
</tr>
<tr>
<td>No</td>
<td>895</td>
<td>−13 (−34 to 8)</td>
</tr>
<tr>
<td>At least one</td>
<td>417</td>
<td>−14 (−49 to 21)</td>
</tr>
</tbody>
</table>

*p Value for interaction. ^Highest quintile.

BMI, body mass index; ETS, environmental tobacco smoke; FEF25–75, forced expiratory flow between 25% and 75% of FVC; FEV1, forced expiratory volume in 1 s; FVC, forced vital capacity.

DISCUSSION

The results of this study suggest that traffic-related air pollution exposure is associated with reduced lung function, especially expiratory flows, in schoolchildren. These results therefore support recent cohort data in the USA demonstrating that exposure to pollution from traffic has adverse effects on...
children’s lung function development. Overall, there was a lack of association between air pollution and respiratory and allergic symptoms which is in agreement with two other studies that used subjective outcome data. However, these results are in contrast with many other investigations documenting significant associations between air pollution and objective measures of asthma and allergic sensitisation.

This is the first study on children’s respiratory health within a community that includes three different methods to assess long-term traffic-related air pollution exposure, including self-reported traffic, measured distance to busy roads and residential NO2 concentrations estimated by a land use regression model. In addition, three different approaches were applied to assess disease status ranging from subjective questionnaire data on allergic and respiratory symptoms to objective measurements of atopy by skin prick tests and lung function by spirometry.

Several findings of our study deserve consideration. We found that the effect on lung function was stronger and more consistent for the two variables that are more closely related to each other, ie, NO2 from land use regression and self-reported traffic intensity. On the other hand, self-reported heavy traffic and distance to busy roads gave more mixed results. It should be recognised that the indicators are complementary as self-reported traffic level and distance to busy roads measured by the geographic information system take other air pollutants (eg, fine and ultrafine particles). Our results are consistent with the findings of the Children’s Health Effects Study in California where distance from motorways was the exposure variable and with the recent results from the Oslo Birth Cohort where NO2, particulate matter PM10 and PM2.5 were considered. We found a similar effect of NO2 on expiratory flow variables that represent peripheral airways, whereas the effect on forced volumes that provide information primarily on damage of the central airways was more limited. Similarly, the recent Swedish birth cohort study found an effect of traffic PM10 and NOx on PEF in 4-year-old children. In addition, our study suggests that the effect of traffic-related air pollution on lung function is not limited to subjects who have asthma, so it is not confined to a well-defined population although there was a tendency for a stronger effect in subjects with atopy.

Some potential weaknesses of this study should be recognised. We used NO2 as an indicator of pollution from traffic because a readily available model for spatial prediction throughout the study region had recently been developed. However, other components from vehicle exhausts that are more or less correlated with NO2 may be responsible for the observed effects. In this study the NO2 concentration estimated by the regression model appears to be correlated with self-reported traffic intensity outside the children’s homes and distance from busy roads. Since exposure was represented by the modelled air pollution levels in the census block where the children lived, some real differences in exposure might have been ignored and consequently the results might be underestimated. In addition, the results were not so much different using traffic intensity as reported by the parents and the measured distance from busy roads as exposure indexes. Although misclassification of true individual exposure levels probably affected the results of this study, the objective assessment of both exposure and disease was done blinded to one another, making such potential bias unlikely to affect the results, except non-differentially.

The present study addressed all schoolchildren in the predefined study area and the high response rate (83.3%) makes selection bias less likely to seriously affect the results. Nevertheless, the study was a cross-sectional investigation and parents were asked to report their children’s lifetime and past year prevalence of allergic and respiratory symptoms. Thus, recall bias and non-differential misclassification of disease from proxy reports may be of concern. On the other hand, this study has several advantages compared with previous investigations, including the three different ways used to assess exposure and

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Figure 2  Associations between different indicators of residential air pollution exposure and respiratory symptoms and sensitisation in children living in Rome, 2000–1, presented as odds ratios (OR) and 95% CI from logistic regression models (adjusted for age, body mass index, father’s education, current mother smoking, indoor humidity or signs of mould during the first year of life and the floor of the subject’s home) calculated for an interquartile range increase in the nitrogen dioxide level (14.16 μg/m3), dichotomous variables of traffic intensity (no–little vs moderate–heavy), heavy traffic (rarely vs sometimes–often) and distance to roads (more or less than 100 m).
disease both objectively and subjectively, as well as a geographically well-defined study area with large exposure contrasts.

A slightly stronger effect of NO\textsubscript{2} on lung function was suggested in girls than boys, which is supported by earlier evidence showing that girls may be more sensitive to air pollution.\textsuperscript{26-33}  

Certainly, contrasting evidence exists from recent cohort data, and there can be several reasons for different effects between studies such as different age ranges more or less close to puberty, different air pollution mixtures, different underlying population susceptibilities and different study designs.\textsuperscript{16} Nevertheless, there is a consistency in the findings of air pollution among girls throughout Europe that calls for further investigation of the underlying relevant mechanism. There was an indication of effect modification by other covariates, suggesting stronger effects in older children, in children with higher maternal education and in children whose parents smoked at home. Although the reasons for these differences are not entirely clear, there are some plausible explanations. Older children may have experienced longer cumulative air pollution exposure on average that has affected their lung growth.\textsuperscript{24} Mother’s education could be a proxy for socioeconomic status, which might point to higher air pollution exposure on average in Rome for those of a higher socioeconomic status living in the central part of the city.\textsuperscript{17} Finally, exposure to parental smoking at home is a well-established detrimental factor for lung function development that could act synergistically with traffic-related air pollution.

Although the biological mechanisms of how traffic-related air pollution reduces lung function in children are not fully understood, several plausible explanations have been discussed in the literature. Air pollution can induce oxidative stress in the respiratory tract resulting in airway inflammation.\textsuperscript{50} Traffic-related air pollution has also been correlated with the amount of carbon in the airway macrophages of children which, in turn, was associated with reductions in lung function measured as FE\textsubscript{V}\textsubscript{1} and forced expiratory flows.\textsuperscript{52} Thus, chronic airway inflammation could produce the decreases in lung function measures like those reported in the present study, and the peripheral airways seem to be mainly affected given the stronger signal that we detected for PEF and FE\textsubscript{F}\textsubscript{50-75}. However, it is difficult to discuss aetiological mechanisms in the context of a cross-sectional study.

The results of this study slightly differed depending on the exposure indicator used. Although air pollution exposure has been assessed as traffic intensity by others,\textsuperscript{4,14} this may be a fairly crude estimate of traffic-related air pollution exposure and especially inappropriate to assess from a questionnaire in relation to subjective symptom reports since subjective measures of both exposure and disease might introduce a correlation of errors. Distance from busy roads appears to be a reasonable proxy for residential air pollution exposure, but it produced different and less consistent results throughout the analyses presented in this study. Despite the weaknesses of the land use regression model, we believe that the predicted NO\textsubscript{2} levels produce the most reliable results and that the analyses of linear regression between these levels and lung function parameters should be regarded as the main results of this study.

In conclusion, the results of this study suggest that residential traffic-related air pollution is associated with decreased expira-
tory lung flows in schoolchildren. However, atopic sensitisation and self-reported respiratory and allergic symptoms were not consistently associated with traffic-related air pollution.

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Ethics approval: The study was approved by the ethical committee of the Catholic University of Rome.

REFERENCES


Environmental exposure


Pulmonary puzzle

Beware the pregnant woman with breathlessness

A 32-year-old woman in the 20th week of her third pregnancy was referred to the advanced lung disease service with breathlessness. As a child she could not keep up with the other children and was told that she had a “hole in the heart” but this did not require further investigation or treatment. In previous pregnancies she had been breathless and fatigued from mid-trimester and into the postpartum period, but she thought this was a normal change in pregnancy.

Examination revealed a soft systolic murmur and loud second heart sound. Echocardiography showed an abnormal dilated left atrium, mildly dilated right ventricle and an estimated pulmonary artery systolic pressure of 60 mm Hg, but no evidence of an atrial septal defect, ventricular septal defect or patent ductus arteriosus. VQ (ventilation–perfusion) scan showed no evidence of thromboembolic disease. High-resolution CT chest excluded interstitial lung disease.

Cardiac MRI (fig 1) revealed an unusual cause for her pulmonary hypertension.

**QUESTION**

What is the cause of her pulmonary hypertension?

*See page 586 for answer.*

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Traffic-related air pollution in relation to respiratory symptoms, allergic sensitisation and lung function in schoolchildren

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