Temporal relationship between air pollutants and hospital admissions for chronic obstructive pulmonary disease in Hong Kong

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Aims: To assess any relationship between the levels of ambient air pollutants and hospital admissions for chronic obstructive pulmonary disease (COPD) in Hong Kong.

Methods: A retrospective ecological study was undertaken. Data of daily emergency hospital admissions to 15 major hospitals in Hong Kong for COPD and indices of air pollutants (sulphur dioxide (SO2), nitrogen dioxide (NO2), ozone (O3), particulates with an aerodynamic diameter of <10 μm (PM10) and 2.5 μm (PM2.5)) and meteorological variables from January 2000 to December 2004 were obtained from several government departments. Analysis was performed using generalised additive models with Poisson distribution, adjusted for the effects of time trend, season, other cyclical factors, temperature and humidity. Autocorrelation and overdispersion were corrected.

Results: Significant associations were found between hospital admissions for COPD with all five air pollutants. Relative risks for admission for every 10 μg/m3 increase in SO2, NO2, O3, PM10 and PM2.5 were 1.007, 1.026, 1.034, 1.024 and 1.031, respectively, at a lag day ranging from lag 0 to cumulative lag 0–5. In a multipollutant model, O3, SO2 and PM2.5 were significantly associated with increased admissions for COPD. SO2, NO2 and O3 had a greater effect on COPD admissions in the cold season (December to March) than during the warm season.

Conclusion: Ambient concentrations of air pollutants have an adverse effect on hospital admissions for COPD in Hong Kong, especially during the winter season. This might be due to indoor exposure to outdoor pollution through open windows as central heating is not required in the mild winter. Measures to improve air quality are urgently needed.

Air pollution is a major public health problem resulting from emissions of air pollutants from both motor vehicles and industrial plants. Previous studies have shown that air pollution is associated with excessive respiratory and cardiovascular morbidity as well as mortality. Hong Kong is a small but densely populated city. The total population was 6.88 million in mid 2004 with a population density of 6380 people per km2. Over the past 10 years the air quality in Hong Kong has worsened substantially. The concentrations of particulates with an aerodynamic diameter <10 μm (PM10) and ozone (O3) recorded at the general air monitoring stations in Hong Kong increased by 15% and 26%, respectively, from 1999 to 2004. The level of nitrogen dioxide (NO2) also increased by 26% between 1991 and 2002. Because of the increase in particulate pollution, impairment in visibility in Hong Kong has also worsened over time. The percentage of the increase in particulate pollution, impairment in visibility and mortality. COPD was the fifth leading cause of death in Hong Kong and accounted for at least 4% of all public hospital acute admissions in 2003. Previous studies have shown that pulmonary function and quality of life are adversely affected by frequent exacerbations, particularly in active smokers. A study conducted in Hong Kong a decade ago showed that air pollution was associated with a significantly increased risk of hospitalisation for patients with respiratory diseases including acute exacerbations of COPD. With increasing air pollution, the effect of air pollutants on acute exacerbations of COPD may be enhanced.

As over 90% of the citizens of Hong Kong are dependent on the public hospital healthcare resources, data on COPD admissions can be retrieved from the central computer system of the Hospital Authority Head Office (HAHO). We assessed the effect of air pollutants—including NO2, sulphur dioxide (SO2), O3, PM10 and particulates with an aerodynamic diameter <2.5 μm (PM2.5)—over a period of 5 years (from January 2000 to December 2004) on emergency admission rates of COPD. Data were collected from 15 major hospitals and 14 air quality monitoring stations in Hong Kong over the same period.

METHODS

This is a retrospective ecological study. Data on emergency hospital admissions for COPD to 15 major hospitals from January 2000 to December 2005 inclusive were obtained from the central computer system of the Hospital Authority Head Office (HAHO). We assessed the effect of air pollutants—including NO2, sulphur dioxide (SO2), O3, PM10 and particulates with an aerodynamic diameter <2.5 μm (PM2.5) —on emergency admission rates of COPD over a period of 5 years (from January 2000 to December 2004) in Hong Kong.

Abbreviations: APHEA2, Air Pollution and Health: A European Approach 2; COPD, chronic obstructive pulmonary disease; NMMAPS, National Morbidity, Mortality, and Air Pollution Study; NO2, nitrogen dioxide; O3, ozone; PM10, particulates with an aerodynamic diameter <10 μm; PM2.5, particulates with an aerodynamic diameter <2.5 μm; SARS, severe acute respiratory syndrome; SO2, sulphur dioxide
the HAHO in Hong Kong. The daily number of hospital admissions with the codes 491, 492 and 496, based on the Ninth Revision of the International Classification of Diseases (ICD-9)\textsuperscript{20} as the primary diagnosis, was recorded.

**Air quality and weather data**

Hourly concentrations of NO\textsubscript{2}, SO\textsubscript{2}, O\textsubscript{3}, PM\textsubscript{10} and PM\textsubscript{2.5} between January 2000 and December 2004 were obtained from the Environmental Protection Department.\textsuperscript{21} The 14 air quality monitoring stations were located in eight districts interspersed in different districts throughout Hong Kong. All the monitoring stations measured levels of NO\textsubscript{2}, SO\textsubscript{2}, O\textsubscript{3} and PM\textsubscript{10}. However, PM\textsubscript{2.5} was measured by three stations only. SO\textsubscript{2} was measured by ultraviolet florescence, NO\textsubscript{2} by chemiluminescence and O\textsubscript{3} by ultraviolet absorption. Both PM\textsubscript{10} and PM\textsubscript{2.5} were measured by the tapered element oscillating microbalance method. The 24 h mean concentrations of NO\textsubscript{2}, SO\textsubscript{2}, PM\textsubscript{10} and PM\textsubscript{2.5} were calculated. Since the formation of O\textsubscript{3} is dependent on sunlight, a daytime (09:00 to 17:00 h) mean concentration was used for analysis. The mean daily temperatures and relative humidity for the same period were obtained from the Hong Kong Observatory.

**Statistical modelling**

The statistical model developed from the Air Pollution and Health: A European Approach 2 (APHEA2) protocol for the time series analyses of mortality and hospital admissions in Europe\textsuperscript{22} was used in this study. A generalised additive model\textsuperscript{23} using a Poisson distribution with log-link function was used to construct a core model. This regressed the daily numbers of COPD admissions on the time variable (day), day of the week variable, daily mean temperature and humidity, and holiday indicator. Smoothing of the time variable was used to control for the long-term seasonal patterns of COPD visits and smoothing splines were used as smoothers. The quasi-likelihood method was used to correct for overdispersion. Each core model was chosen based on the degree of freedom that gave the minimum Akaike’s information criterion (AIC) value.\textsuperscript{24}

Autocorrelation was adjusted by adding autoregressive terms to the model. After controlling for the confounding effects of seasonality, days of the week and climatic variables, daily concentrations of PM\textsubscript{10}, NO\textsubscript{2}, SO\textsubscript{2} and PM\textsubscript{2.5} were added to the core model to determine the relative risk (RR) of COPD admissions for a 10 μg/m\textsuperscript{3} increase in each of these air pollutants. Concentrations of all the air pollutants for the same day (lag 0) up to lag days (lag 5) and cumulative lags by 2 (average of lag 0 and 1), 3 (average of lag 0, 1 and 2) to 6 days (average of lag 0–5) were tested in each model. The lag day with the air pollutant concentration that yielded the largest χ\textsuperscript{2} value obtained from the change in deviances\textsuperscript{25} was chosen.

Multipollutant models were run for air pollutants that were significant in the single pollutant analysis, and the lag with the strongest univariate effect was tested. The correlation coefficients of the individual air pollutants were examined and only those with r>0.7 were entered into the model. Pairwise analyses were performed by entering each pollutant, cold season indicator and their interaction term into the core model. December to March was considered to be the cold season in Hong Kong with a mean daily temperature below 20°C.\textsuperscript{5} The standard errors of the estimates were computed using the supplementary program used in the re-analysis of the National Morbidity, Mortality, and Air Pollution Study (NMMAPS).\textsuperscript{25} All calculations were performed with the software S-plus 4.0 using a more stringent convergence criteria in the gam(.) function to prevent the convergence problem.\textsuperscript{26}

**RESULTS**

There were 119,225 admissions for acute exacerbations of COPD from January 2000 to December 2004 with mean (SD) annual and daily admissions of 23,845 (3645) and 65.3 (18.9), respectively. Figure 1 shows a plot of daily admissions over the 5 year period with the Pearson residuals. The seasonal or long-term pattern was not apparent in the residuals.

The means, standard deviations, ranges and percentiles of the daily counts of hospital admissions, daily levels of NO\textsubscript{2}, SO\textsubscript{2}, PM\textsubscript{10} and PM\textsubscript{2.5} and the levels of O\textsubscript{3} from 09:00 to 17:00 h are shown in table 1. Correlations of the pollutants are shown in table 2; a close correlation was found between PM\textsubscript{10} and PM\textsubscript{2.5} (r = 0.95, p<0.001).

There was a statistically significant association between hospital admissions for acute exacerbations of COPD and an increase in 10 μg/m\textsuperscript{3} of all five pollutants (NO\textsubscript{2}, PM\textsubscript{10}, O\textsubscript{3}, SO\textsubscript{2} and PM\textsubscript{2.5}; table 3). A lag effect (cumulative from day 0 to ≥2 days) was found for all the pollutants assessed except SO\textsubscript{2}. The strongest effect on COPD admissions observed for SO\textsubscript{2} (RR 1.007, 95% CI 1.001 to 1.014) was on lag 0 (ie, no lag days), whereas for NO\textsubscript{2} (RR 1.026, 95% CI 1.022 to 1.031) it was on cumulative lag days 0–3 as indicated by the greatest χ\textsuperscript{2} score (“best” lag). A more delayed effect on COPD admission was observed for O\textsubscript{3} (RR 1.034, 95% CI 1.030 to 1.040), PM\textsubscript{10} (RR 1.024, 95% CI 1.021 to 1.028) and PM\textsubscript{2.5} (RR 1.031, 95% CI 1.026 to 1.036), with the “best” cumulative lag days of 0–5. The overdispersion parameter (φ) and the autocorrelation coefficients were 1.295 and −0.02163, respectively. The intercorrelation between the coefficients (β) of the pollutants in the
multim pollutant model, based on the covariance matrix,\textsuperscript{29} ranged from $-0.4161$ to $0.2005$.

We noted a sharp fall in the number of daily admissions in April 2003, which was probably due to the major outbreak of severe acute respiratory syndrome (SARS) in Hong Kong at that time.\textsuperscript{27}–\textsuperscript{28} As a sensitivity analysis, we constructed separate models for hospital admissions in 2000–2002, 2003 and 2004 by analysing the data before, during and after the SARS epidemic in 2003. The RRs before 2003 were lower than those after 2003 for NO\textsubscript{2}, O\textsubscript{3}, PM\textsubscript{10} and PM\textsubscript{2.5}, but the results were largely similar to those when the entire period was used for the model (data available in supplemental table 1 available online at www.thorax.bmj.com/supplemental).

The multim pollutant model was constructed and the results are shown in table 4. SO\textsubscript{2}, O\textsubscript{3} and PM\textsubscript{2.5} were statistically most significantly related to COPD admissions. Of all the pollutants, O\textsubscript{3} had the highest $\chi^2$ score, indicating that it contributed to the greatest risk of admission for COPD (RR 1.029, 95% CI 1.022 to 1.036). There was only moderate correlation ($r<0.5$) between most air pollutants, except for the correlations between NO\textsubscript{2} and SO\textsubscript{2} ($r = 0.66$) and between PM\textsubscript{2.5} and PM\textsubscript{10} ($r>0.9$). Since PM\textsubscript{2.5} is widely believed to be more damaging to health and has a higher RR for admission to hospital with an acute exacerbation of COPD than PM\textsubscript{10}, the former was selected into the statistical model instead.

There was also a seasonal effect on the air pollutants. Using 20\textdegree C as the cut-off temperature, a lower temperature was associated with a higher level of mean daily PM\textsubscript{10}, PM\textsubscript{2.5} and NO\textsubscript{2} but a lower level of mean 8 h O\textsubscript{3}. In addition, the mean number of daily admissions was higher with a lower temperature (74.5 vs 61.9, $p<0.0001$). The interaction between seasons and individual pollutants is shown in table 5. An additive effect of the cold weather on the pollutants was observed on COPD admissions for NO\textsubscript{2}, O\textsubscript{3} and SO\textsubscript{2} (RR 1.016 (95% CI 1.009 to 1.024), 1.020 (95% CI 1.008 to 1.031) and 1.0201 (95% CI 1.008 to 1.033), respectively).

DISCUSSION

This study assessed the effects of air pollution specifically on hospital admissions for acute exacerbations of COPD. All five air pollutants (NO\textsubscript{2}, SO\textsubscript{2}, O\textsubscript{3}, PM\textsubscript{10} and PM\textsubscript{2.5}) had a positive association with hospital admissions for acute exacerbations of COPD, with different magnitudes in terms of the relative risks and the number of lag days. Previous time series studies in Asia did not focus on COPD alone but on admissions for all respiratory diseases or all-cause mortality.\textsuperscript{51, 10129} As the time frame chosen in this study was recent and for an extended duration (from January 2000 to December 2004), this study provides up to date information on the effects of air pollutants specifically on COPD.

The total number of hospital admissions in this single-city study was high ($n = 119,225$) when compared with other US national multi-city studies; for example, a US multi-city study by Medina-Ramon et al included a total of 578,006 admissions.\textsuperscript{2} As there was a sharp decline in the number of daily admissions in April 2003 which was most probably caused by the major outbreak of SARS,\textsuperscript{27–28} we constructed separate models to analyse hospital admissions and air pollution data for the periods before, during and after the SARS epidemic in 2003. Although the RRs before 2003 were lower than those after 2003 for NO\textsubscript{2}, O\textsubscript{3}, PM\textsubscript{10} and PM\textsubscript{2.5}, the results were similar to those when the entire period was used for the model. Not only has this confirmed the robustness of the general additive model in correcting for the potential effects caused by an epidemic that might affect the number of hospital admissions, it has shown that the relationship between air pollution and COPD admissions was not affected by an event that reduced hospital admissions but not air pollution.

### Table 1 Daily summary of the number of COPD admissions, meteorological and pollutant data

<table>
<thead>
<tr>
<th></th>
<th>Mean</th>
<th>SD</th>
<th>Min</th>
<th>Q1</th>
<th>Median</th>
<th>Q3</th>
<th>Max</th>
</tr>
</thead>
<tbody>
<tr>
<td>Admissions for COPD (n)</td>
<td>65.3</td>
<td>18.9</td>
<td>15</td>
<td>52</td>
<td>65</td>
<td>78</td>
<td>125</td>
</tr>
<tr>
<td>Temperature (\textdegree C)</td>
<td>23.7</td>
<td>4.9</td>
<td>8.2</td>
<td>19.9</td>
<td>25.2</td>
<td>27.9</td>
<td>31.0</td>
</tr>
<tr>
<td>Relative humidity (%)</td>
<td>74.4</td>
<td>12.8</td>
<td>21.1</td>
<td>67.6</td>
<td>77.0</td>
<td>83.0</td>
<td>97.0</td>
</tr>
<tr>
<td>NO\textsubscript{2} (\mu g/m\textsuperscript{3})</td>
<td>51.2</td>
<td>21.8</td>
<td>10.1</td>
<td>35.0</td>
<td>50.1</td>
<td>64.0</td>
<td>157.6</td>
</tr>
<tr>
<td>PM\textsubscript{10} (\mu g/m\textsuperscript{3})</td>
<td>50.1</td>
<td>23.9</td>
<td>13.6</td>
<td>31.9</td>
<td>44.5</td>
<td>64.1</td>
<td>172.2</td>
</tr>
<tr>
<td>O\textsubscript{3} (\mu g/m\textsuperscript{3})</td>
<td>31.0</td>
<td>16.2</td>
<td>4.8</td>
<td>18.1</td>
<td>26.7</td>
<td>41.0</td>
<td>101.3</td>
</tr>
<tr>
<td>SO\textsubscript{2} (\mu g/m\textsuperscript{3})</td>
<td>15.0</td>
<td>11.6</td>
<td>0.7</td>
<td>7.6</td>
<td>12.5</td>
<td>19.5</td>
<td>99.4</td>
</tr>
<tr>
<td>PM\textsubscript{2.5} (\mu g/m\textsuperscript{3})</td>
<td>35.7</td>
<td>20.6</td>
<td>6.0</td>
<td>19.4</td>
<td>31.7</td>
<td>46.7</td>
<td>163.2</td>
</tr>
</tbody>
</table>

SD, standard deviation; Q1, 25th percentile; Q3, 75th percentile; Min, minimum; Max, maximum; COPD, chronic obstructive pulmonary disease; NO\textsubscript{2}, nitrogen dioxide; O\textsubscript{3}, ozone; PM\textsubscript{10}, PM\textsubscript{2.5}, particulates with an aerodynamic diameter <10 \mu m or <2.5 \mu m; SO\textsubscript{2}, sulphur dioxide.

### Table 2 Correlations between different air pollutants and meteorological variables

<table>
<thead>
<tr>
<th></th>
<th>SO\textsubscript{2}</th>
<th>NO\textsubscript{2}</th>
<th>PM\textsubscript{10}</th>
<th>O\textsubscript{3}</th>
<th>PM\textsubscript{2.5}</th>
<th>Temperature</th>
<th>Humidity</th>
</tr>
</thead>
<tbody>
<tr>
<td>SO\textsubscript{2}</td>
<td>r = 1.00*</td>
<td>0.659*</td>
<td>0.229*</td>
<td>-0.037*</td>
<td>0.282</td>
<td>0.084*</td>
<td>0.169*</td>
</tr>
<tr>
<td>NO\textsubscript{2}</td>
<td>r = 1.00</td>
<td>-0.001</td>
<td>-0.001</td>
<td>0.111</td>
<td>&lt;0.001</td>
<td>&lt;0.001</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>PM\textsubscript{10}</td>
<td>r = 1.00</td>
<td>0.421*</td>
<td>0.952</td>
<td>-0.344*</td>
<td>-0.390*</td>
<td>&lt;0.001</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>O\textsubscript{3}</td>
<td>r = 1.00</td>
<td>0.394</td>
<td>0.236</td>
<td>0.268*</td>
<td>0.212*</td>
<td>&lt;0.001</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>PM\textsubscript{2.5}</td>
<td>r = 1.00</td>
<td>-0.356</td>
<td>-0.343</td>
<td>-0.001</td>
<td>&lt;0.001</td>
<td>1.00</td>
<td>0.127*</td>
</tr>
<tr>
<td>Temperature</td>
<td>r = 1.00</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
</tr>
<tr>
<td>Humidity</td>
<td>r = 1.00</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
</tr>
</tbody>
</table>

r, Pearson correlation coefficient; SO\textsubscript{2}, sulphur dioxide; NO\textsubscript{2}, nitrogen dioxide; O\textsubscript{3}, ozone; PM\textsubscript{10}, PM\textsubscript{2.5}, particulates with an aerodynamic diameter <10 \mu m or <2.5 \mu m.
The concentrations of PM$_{10}$ and PM$_{2.5}$ in Hong Kong were high in 2000–4 compared with other western countries. The mean (SD) level of PM$_{10}$ was 50.1 (23.9) g/m$^3$ in the current study, compared with a level of 30.4 (5.1) g/m$^3$ in a study involving 36 American cities in 1986–99. Our median level of PM$_{10}$ (44.5 g/m$^3$) was also higher than those in several European studies (median levels for Stockholm, Birmingham, London and Netherlands 13.6, 21.5, 24.9 and 33.4 g/m$^3$, respectively). For PM$_{2.5}$, the median (IQR) level in this current study was 31.7 (19.4–46.7) g/m$^3$, which was much higher than the level of 13.4 (11.3–15.2) g/m$^3$ in a national US study.

Our study had comparable levels of SO$_2$, O$_3$ and NO$_2$ to those in western European cities such as Rome. We found that a 10 g/m$^3$ increase in PM$_{10}$ was associated with a 2.4% increase in hospital admissions for acute exacerbations of COPD at cumulative lag days of 0–5. The effect of PM$_{10}$ on respiratory admissions appeared to be less than in other studies. A 10 g/m$^3$ increase in PM$_{2.5}$ was associated with a 3.1% increase in COPD admissions with a cumulative lag of 0–5 in our study using the APHEA2 statistical model, in contrast to a recent US study using the NMMAPS statistical model which found the largest effect per 10 g/m$^3$ increase in PM$_{2.5}$ occurring at lag 0 and 1 with a risk of about 0.9%. Analysis of our data using the NMMAPS model has shown that each 10 g/m$^3$ increase in PM$_{2.5}$ was associated with a 2.0% increase in COPD admissions, and the effect of PM$_{2.5}$ was higher than that reported by Dominici et al. The higher impact of the PM$_{2.5}$ level on respiratory admissions as noted in our study may be due to a higher level of PM$_{2.5}$ in our community and a dose-response effect of the pollutants on respiratory morbidity. The effect of particulate matter on respiratory morbidity remains a focus of research. The leading hypothesis emphasised inflammatory responses in the lungs and release of cytokines with local and systemic consequences. Particulate matters may promote inflammation, thereby exacerbating underlying lung diseases and reducing the efficacy of lung defence mechanisms.

Despite several previous reports from Hong Kong on increasing air pollution, the average values detected by the air quality monitoring stations for NO$_2$, SO$_2$, O$_3$ and PM$_{10}$ did not differ significantly from a previous study conducted by Wong et al. in Hong Kong in 1994–5 on the effect of air pollutants on all respiratory admissions. In fact, the effect of individual pollutants on COPD in our study was similar to that in the study by Wong et al.

### Table 3: Relative risk (with 95% CI) for the pollutants per 10 μg/m$^3$ increase in the concentration of air pollutants for hospitalisations due to an acute exacerbation of COPD (single pollutant model)

<table>
<thead>
<tr>
<th>Lag days</th>
<th>NO$_2$</th>
<th>PM$_{10}$</th>
<th>O$_3$ (8 h)</th>
<th>SO$_2$</th>
<th>PM$_{2.5}$</th>
</tr>
</thead>
<tbody>
<tr>
<td>Lag 0</td>
<td>1.009 (1.005 to 1.013)**</td>
<td>1.003 (1.000 to 1.005)*</td>
<td>1.002 (0.998 to 1.005)</td>
<td>1.007 (1.001 to 1.014)†</td>
<td>1.002 (0.998 to 1.001)</td>
</tr>
<tr>
<td>Lag 1</td>
<td>1.001 (0.997 to 1.005)</td>
<td>1.005 (1.002 to 1.007)**</td>
<td>1.011 (1.008 to 1.015)**</td>
<td>0.991 (0.981 to 1.001)</td>
<td>1.003 (0.999 to 1.007)**</td>
</tr>
<tr>
<td>Lag 2</td>
<td>1.003 (0.999 to 1.007)</td>
<td>1.010 (1.007 to 1.012)**</td>
<td>1.011 (1.008 to 1.020)**</td>
<td>0.992 (0.985 to 1.000)</td>
<td>1.011 (1.007 to 1.014)**</td>
</tr>
<tr>
<td>Lag 3</td>
<td>1.010 (1.007 to 1.014)**</td>
<td>1.008 (1.006 to 1.011)**</td>
<td>1.011 (1.008 to 1.015)**</td>
<td>1.004 (0.998 to 1.011)</td>
<td>1.011 (1.008 to 1.015)**</td>
</tr>
<tr>
<td>Lag 4</td>
<td>1.008 (1.004 to 1.012)**</td>
<td>1.007 (1.004 to 1.009)**</td>
<td>1.006 (1.003 to 1.010)**</td>
<td>1.004 (0.997 to 1.010)</td>
<td>1.009 (1.006 to 1.013)**</td>
</tr>
<tr>
<td>Lag 0–1</td>
<td>1.007 (1.003 to 1.011)**</td>
<td>1.005 (1.002 to 1.008)**</td>
<td>1.011 (1.006 to 1.014)**</td>
<td>0.998 (0.991 to 1.006)</td>
<td>1.004 (0.999 to 1.008)**</td>
</tr>
<tr>
<td>Lag 0–2</td>
<td>1.009 (1.004 to 1.013)**</td>
<td>1.011 (1.008 to 1.014)**</td>
<td>1.019 (1.015 to 1.023)**</td>
<td>0.993 (0.985 to 1.001)</td>
<td>1.010 (1.006 to 1.015)**</td>
</tr>
<tr>
<td>Lag 0–3</td>
<td>1.026 (1.022 to 1.031)**</td>
<td>1.016 (1.013 to 1.019)**</td>
<td>1.027 (1.022 to 1.031)**</td>
<td>0.998 (0.989 to 1.007)</td>
<td>1.018 (1.013 to 1.022)**</td>
</tr>
<tr>
<td>Lag 0–4</td>
<td>1.021 (1.017 to 1.026)**</td>
<td>1.020 (1.017 to 1.024)</td>
<td>1.031 (1.027 to 1.036)**</td>
<td>1.001 (0.991 to 1.010)</td>
<td>1.024 (1.019 to 1.029)**</td>
</tr>
<tr>
<td>Lag 0–5</td>
<td>1.026 (1.022 to 1.031)**</td>
<td>1.024 (1.021 to 1.028)**†</td>
<td>1.034 (1.030 to 1.040)**†</td>
<td>1.004 (0.994 to 1.014)</td>
<td>1.031 (1.026 to 1.036)**†</td>
</tr>
</tbody>
</table>

SO$_2$, sulphur dioxide; NO$_2$, nitrogen dioxide; O$_3$, ozone; PM$_{10}$, PM$_{2.5}$, particulates with an aerodynamic diameter <10 μm or <2.5 μm. *p<0.05, **p<0.001. †Highest χ² score (best lag).

### Table 4: Relative risk (with 95% CI) for the pollutants per 10 μg/m$^3$ increase in the concentration of air pollutants for hospitalisations due to an acute exacerbation of COPD (mult pollutant model)

<table>
<thead>
<tr>
<th>NO$_2$ (lag 0–3)</th>
<th>O$_3$ (8 h) (lag 0–5)</th>
<th>SO$_2$ (lag 0)</th>
<th>PM$_{2.5}$ (lag 0–5)</th>
</tr>
</thead>
<tbody>
<tr>
<td>0.993 (0.985 to 1.001)</td>
<td>1.029 (1.022 to 1.036)**†</td>
<td>1.008 (1.001 to 1.015)*</td>
<td>1.014 (1.007 to 1.022)*</td>
</tr>
<tr>
<td>1.029 (1.022 to 1.034)**†</td>
<td>1.008 (1.001 to 1.015)*</td>
<td>1.011 (1.004 to 1.017)*</td>
<td></td>
</tr>
</tbody>
</table>

SO$_2$, sulphur dioxide; NO$_2$, nitrogen dioxide; O$_3$, ozone; PM$_{10}$, PM$_{2.5}$, particulates with an aerodynamic diameter <10 μm or <2.5 μm. *p<0.05. †Highest χ² score in the pollutant model.
NO\textsubscript{2}, SO\textsubscript{2} and PM\textsubscript{10} was assessed only for up to 4 days (lag 0–3) by Wong et al, whereas we have extended the cumulative lag to 6 days (lag 0–5) for all the pollutants. We noted the most statistically significant effect of PM\textsubscript{10} on COPD admissions was at lag 0–5, in contrast to lag 0–3 in the study by Wong et al.\textsuperscript{3} Concerning the use of cumulative lags in analysing the effects of pollutants on admissions for COPD, the RRs derived from two different averaging period variables must be interpreted with caution. For example, it requires a longer duration of exposure to effect a change of 1 \textmu g/m\textsuperscript{3} in a 5 day average (ie, a 1 \textmu g/m\textsuperscript{3} increase on 5 consecutive days) than for a 1 day average increase of 1 \textmu g/m\textsuperscript{3} (ie, a 1 day increase of 1 \textmu g/m\textsuperscript{3} on a single day only). Our study has provided additional information on the effect of PM\textsubscript{2.5} on COPD admissions which was not addressed in the previous study.\textsuperscript{2} Furthermore, a multipurpouse model specific for acute exacerbations of COPD was not constructed in the study by Wong et al.\textsuperscript{3} In our study, when a number of pollutants (NO\textsubscript{2}, SO\textsubscript{2}, O\textsubscript{3} and PM\textsubscript{2.5}) were analysed together, all except NO\textsubscript{2} were retained in the model. As expected, the RRs of the individual pollutants were smaller than those in the single pollutant models, with O\textsubscript{3} having the greatest effect on COPD admissions.

We found differences in the maximum effect of the individual air pollutants on COPD hospitalisations in terms of the number of cumulative lag days. Despite the use of the “best” lag calculated according to statistical criteria, these differences could possibly be explained by the chemical and toxicological properties of the individual pollutants on the airway. SO\textsubscript{2} is very soluble in the upper respiratory tract\textsuperscript{7} and thus may produce an immediate irritant effect on the respiratory mucosa which would account for the fact that no lag days were observed for SO\textsubscript{2}. In contrast, O\textsubscript{3} and NO\textsubscript{2} are much less soluble, enabling them to penetrate deeply into the lungs. In addition, these two gaseous pollutants and PM\textsubscript{2.5} are highly reactive oxidants and can cause inflammation of the respiratory epithelium at high concentrations.\textsuperscript{35,36} An experimental chamber study reported a decrease in lung function after exposure to O\textsubscript{3}.\textsuperscript{37} On the other hand, exposure to NO\textsubscript{2} might enhance the recruitment of macrophages and T lymphocytes to the airway, as shown by increased CD45RO+ lymphocytes, B cells and natural killer (NK) cells in the bronchoalveolar lavage fluid of healthy volunteers.\textsuperscript{40} As oxidant-induced airway inflammation takes time to develop, this may account for some cumulative lag days with O\textsubscript{3}, NO\textsubscript{2} and PM\textsubscript{10} rather than an immediate effect on admissions for acute exacerbations of COPD.

In contrast to this current study, previous studies found an increased risk of admissions for respiratory diseases, including COPD and pneumonia, associated with ambient O\textsubscript{3} and PM\textsubscript{10} levels predominantly in the warm season.\textsuperscript{2,41,42} A previous study in the US suggested that the difference in the positive effect of O\textsubscript{3} on respiratory admissions in different seasons might be due to higher O\textsubscript{3} concentrations and more time spent outdoors during the warm season. Since O\textsubscript{3} is a highly reactive gas, its indoor concentration is extremely low in buildings with closed windows.\textsuperscript{2} In Hong Kong the situation differs from western countries as the winter season is not cold enough to require central heating in most places. Windows are often opened in winter for ventilation purposes. Summers are generally hot and humid, and most people spend more time indoors with air conditioning. This might explain the enhanced effects of O\textsubscript{3} exposure—and probably SO\textsubscript{2} and NO\textsubscript{2} exposure—in patients with COPD in the winter in HK. This is supported by a previous study showing that air conditioned homes had lower air exchange rates than homes with windows opened for ventilation.\textsuperscript{43} The reason for the seasonal differences observed in our current study compared with other western studies is not entirely clear and requires further study.

This study has a few limitations. As with other ecological studies, it was limited by the fact that precise exposure to a specific pollutant could not be assessed. Also, the ICD coding for identifying COPD hospital admissions might be limited by the accuracy of coding. In addition, the interaction between pollutants with the collinearity issues and limitations of the time series design made it difficult to assess which particular pollutant had a direct adverse effect on the patients.

In summary, air pollutant levels of SO\textsubscript{2}, NO\textsubscript{2}, O\textsubscript{3}, PM\textsubscript{10} and PM\textsubscript{2.5} were all associated with increased hospital admissions for acute exacerbations of COPD in Hong Kong. O\textsubscript{3} was the most important air pollutant leading to increased hospitalisations for acute exacerbations of COPD, and PM\textsubscript{10} and PM\textsubscript{2.5} levels in Hong Kong were higher than in other countries. The PM\textsubscript{10} concentration exceeded the European Union recommendation (average 24 h concentration <40 \textmu g/m\textsuperscript{3} by 2005).\textsuperscript{12} Public health measures are urgently needed to improve air quality in Hong Kong.

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Further data are given in the online supplement available at www.thorax.bmj.com/supplemental

Table 5 Interaction between season and individual pollutants

<table>
<thead>
<tr>
<th>Pollutant</th>
<th>RR</th>
<th>95% CI</th>
<th>p Value</th>
<th>t Score</th>
</tr>
</thead>
<tbody>
<tr>
<td>NO\textsubscript{2} (lag 0–3)</td>
<td>1.0164</td>
<td>1.0086 to 1.0243</td>
<td>&lt;0.0001*</td>
<td>4.150476</td>
</tr>
<tr>
<td>O\textsubscript{3} (lag 0–5)</td>
<td>1.0195</td>
<td>1.0081 to 1.0309</td>
<td>0.0007*</td>
<td>3.379242</td>
</tr>
<tr>
<td>SO\textsubscript{2} (lag 0)</td>
<td>1.0205</td>
<td>1.0078 to 1.0334</td>
<td>0.0015*</td>
<td>3.178459</td>
</tr>
<tr>
<td>PM\textsubscript{2.5} (lag 0–5)</td>
<td>0.9983</td>
<td>0.9891 to 1.0076</td>
<td>0.5200</td>
<td>0.643313</td>
</tr>
<tr>
<td>PM\textsubscript{10} (lag 0–3)</td>
<td>1.0012</td>
<td>0.9951 to 1.0073</td>
<td>0.7059</td>
<td>0.37738</td>
</tr>
</tbody>
</table>

RR, relative risk (the warm season was used as the reference). The cold season is defined as the period from December to March of each year.

*p<0.05.
REFERENCES


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